

AEROBIC BACTERIAL DEGRADATION OF KITCHEN WASTE: A REVIEW

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ABSTRACT

Kitchen waste contain easily degradable an organic biomolecules substances as feedstock for aerobic bio-reactor which works as aerobic digester system to produce highly simple organic and inorganic matters containing biomass as biofertilizers. To create an excellent biomass used as biofertilizers source which will be more valuable and an effective, eco-friendly, cut down on landfill waste, generates a high-quality biomass and reduces CH₄, CO₂ emissions & controlled global warming effect. Therefore these bio-fertilizers contain beneficial & active bacterial communities which involved in the regulation of soil properties on the basis of their biological activity. In the presence of oxygen, aerobic microbes play a major role in the degradation of complex organic materials into mainly to simple biomass & CO₂ production. Aerobic treatment has favourable effects like removal of higher organic concentration, various pathogenic bacteria removal and also produces a stable biomass. The bio-digester requires addition of sodium hydroxide (NaOH) to maintain the alkalinity and pH to 7. For this reactor we have prepared an excellent bacterial community which applied into mixture of kitchen waste slurry along with other biowaste in aerobic bioreactor for biomass as humus production in large quantity and therefore, may be controlled of environmental pollution. A combination of these mixed an excellent bacterial community is used for biomass production at different temperature in laboratory.

Keywords: Aerobic degradation, Waste Treatment, Biomass as biofertilizers & CO₂

INTRODUCTION

India is rich agricultural resources, accounts for 50 million tons of vegetable waste, which is about 30 % of its total production (Verma *et al.*, 2011). Food waste is an unwanted raw or cooked food throw away during or after food preparation that is no longer fit for consumption or desirable (Jean *et al.*, 2009). As food waste has high moisture content and can be decomposed rapidly, many unpleasant environmental consequences can arise during its storage, collection and transportation (Choi and Park, 1998). Microbial population of activated sludge is capable of degrading simple chemical structures from municipal wastewater, e.g. carbohydrates, short-chain fatty acids, simple alcohols, proteins and amino acids (Pike, 1975). Management and removal of such pollutants are

facing a crucial state, due to the unavailability of suitable strategies for the treatment and waste disposal. Unscientific disposal can cause an adverse impact on all components of the environment and human health (Rathi, 2006; Gupta *et al.*, 1998).

Asian countries alone spent about 25 billion US dollars on solid waste management during the year 1990, which was expected to rise up to 50 billion US dollars per annum by 2025 (Hoorweg and Thomas, 1999). About 135.5 million tones/year of municipal solid waste is generated in India and it is consist of 30-40% of food waste (Kashyap *et al.*, 2003). Human societies generate large amounts of waste (Taylor *et al.*, 2002).

Table 1 Total production of municipal solid waste (MSW), industrial and agricultural waste in million tons/year. (Source: Kashyap *et al.*, 2003)

Waste	India	Brazil	Sudan	USA	Sweden
Municipal Solid Waste	135.5	44.0	2.3	148.0	5.3
Sewage	44.9	8.02	1.4	16.0	0.6
Manure	653.0	470.0	68.0	306.0	13.2
Agricultural residues	200.0	47.0	8.1	573.0	12.6
Biomass	140.0	496.8	192.3	427.0	14.0

Although organic matter can also be decomposed under anaerobic condition, the degradation is slow and less efficient, and produces less heat and more undesirable products, including CH₄ and N₂O, which are greenhouse gases contributing to global warming (Hao *et al.*, 2001).

In comparison, aerobic composting is accepted as an eco-friendly an option for handling food waste, because the predominant aerobic environment can help mitigate methane generation. Thus, composting is presently gaining more and more attention in treating food waste (Droffner and Brinton, 1995; Elwell *et al.*, 1996; Donahue *et al.*, 1998; Laos *et al.*, 1998; Faucette *et al.*, 2001; Seymour *et al.*, 2001; Tomati *et al.*, 2001; Filippi *et al.*, 2002; Das *et al.*, 2003; Cekmecelioglu *et al.*, 2005; Chang *et al.*, 2005a). Pathogenic microorganisms and harmful chemicals in solid waste can be contaminated into the environment when the waste is not properly managed (Wai-Ogosu, 2004; Ogbonna and Igbenije, 2006). Waste can contaminate surface water, ground water, soil and air

which cause more problems for humans, other species, and ecosystems (Obire *et al.*, 2002).

KITCHEN WASTE SOURCES

In Municipal Solid Waste, food and organic waste mostly consist of uneaten food and food waste generated to especially from residences (School, hospitals, universities, offices), and commercial (from restaurants, cafeteria, hotels, markets and industry) sources known as kitchen waste (Village, 1998).

COMPOSITION OF KITCHEN WASTE

Kitchen waste is characterized by a high organic content containing soluble sugars, starch, lipids, proteins, cellulose, and other compounds that are readily biodegradable, and generally contain few compounds that inhibit bacteria (Wang *et al.*, 2003b).

MICROORGANISMS INVOLVED

Macdonald *et al.* (1981) noted that the composting process is brought about by several organisms such as bacteria, fungi, actinomycetes and protozoa and may also involve invertebrates such as nematodes, pot worms, earthworms, mites and various other organisms.

BIODEGRADATION OF KITCHEN WASTE

Composting is considered one of the most suitable approaches for disposing of solid waste and for increasing the amount of organic matter that can be used to restore and preserve the environment (Stentiford, 1987). Aerobic composting involves a process of biological decomposition and stabilization of organic substrates under conditions that allow multiplication and activity of the thermophilic microorganisms as a result of biologically produced heat, to produce a final product that is stable, free of pathogens, pests and plant seeds, useful in agriculture and forestry as manure (Balasundaran *et al.*, 1999; Saravanan *et al.*, 2003).

Donahue *et al.* (1998) demonstrated a successful composting of food waste with sawdust and mulch chip in an in-vessel system within 14 days. Composting reduced the amount of disposed garbage, recycled the organic resources and produced the organic fertilizer for organic farming. However, the variation of compost quality is huge due to differences of raw materials, operation conditions, micro flora, and composting methods (Aoshima *et al.*, 2001; Tsai *et al.*, 2002). There will be an advantage if the kitchen waste with high fraction of organic content can be utilized as a high value of carbon resource. So far, the recovery energy from methane fermentation and the production of organic fertilizers by composting using kitchen waste has been implemented (Wang *et al.*, 2001). These organic wastes also have been used for organic acids production. An organic acids production from kitchen waste, not only can eliminate waste pollution problem but also reduce the production cost of organic acids (Wang *et al.*, 2003b). Lactic acid could be stably accumulated during kitchen waste fermentation by controlling some fermentation parameters such as temperature, pH etc (Sakai *et al.*, 2000; Wang *et al.*, 2002). Lactic acid was found to be the main fermentation products for kitchen wastes (Wang *et al.*, 2001; Wang *et al.*, 2003b).

In recent years, composting has attracted much attention and has come to be regarded as an environmentally friendly and sustainable alternative for the management and recycling of organic wastes (Saebo and Ferrini, 2006). Composting is being widely employed for the treatment of sewage sludge, yard waste, food waste, forestry industry waste, municipal solid waste, and agricultural waste (Liang, 2000). Composting is a thermogenic, solid state fermentation process, carried out by a succession of microbial populations beginning with mesophilic bacteria, actinomycetes and fungi followed by thermophiles and ending again with mesophiles (Johri *et al.*, 1999). Composting process creates stable, soil-enriching humus and concentrates the Nitrogen (N), Phosphorous (P), Potassium (K), Calcium (Ca) and Magnesium (Mg) contents (Eneji *et al.*, 2001). Composting is useful in avoiding green house gas (GHG) emissions, as it is an aerobic process. Composting with sufficient aeration generates biogenic CO₂ instead of CH₄ from the degradation of organic materials. It also reduces emissions from organic waste by a significant amount, compared to land filling without energy recovery (1.2 t CO₂ e/t for food wastes and 0.7 t CO₂ e/t for yard trimmings) (Mohareb *et al.*, 2004). Moreover, composting produces a useful byproduct that can be used as a soil conditioner.

Composting converts various components in organic wastes into relatively stable substances that can be used as a soil amendment or organic fertilizer. However, composting can also impact negatively on the environment, namely through the generation of odorous gaseous emissions (Pagans *et al.*, 2006). Consequently, the problems of odor emission in composting plants require immediate attention if composting is to become a viable option for the industrial scale recycling of food waste. More than 100 kinds of odorous gases are emitted from composting processes, of which the nitrogen-containing compounds, sulfur-containing compounds, and short-chain fatty acids have attracted the most attention due to their low threshold limits (Curties, 1981).

According to Verrier *et al.*, (1987) and Raynal *et al.*, (1998) the organic fraction of the waste includes about 75% sugars and hemicelluloses, 9% cellulose, and 5% lignin, carbohydrates, amino acids, fatty acids and their esters. The use of bioremediation technologies for removing these contaminants provides a safe and economic alternative to commonly used physical-chemical treatment. The exploitation of the metabolic versatility of microorganisms is advantageous in bioremediation but the actual number of degraders of a target compound may only 5-10% of the total microbial community (Chandrakant and Rao, 2011). Vegetable waste provides good amount of nutrients for inhabiting microbes, they are neither pathogens nor concerned with human health. However they are prone to strong odors during decomposition. The high moisture contain of vegetable waste causes expensive to dispose it off. The vegetable waste containing lettuce and onion waste has high amounts of sulfur 0.2% and 0.7% respectively and their moisture content is of 96.2% and 91.1% respectively (McGuckin *et al.*, 1999).

Bio-fertilizer are commonly contains living microorganisms and their activities will promote soil ecosystem and produce supplementary substance for the plants (Parr *et al.*, 2002). The microorganisms and the nutrients occur in the raw materials which are helpful in improving of soil health. There are different types of bio-fertilizers available that their differences are mainly the raw materials used, forms of utilization and the sources of microorganisms (Svensson *et al.*, 2004).

ROLE OF MICROBIAL ENZYMES IN COMPOSTING

Bacterial extracellular enzyme mediated activity is the major process involved in the hydrolysis of organic pollutants. Bacterial communities contain a broad range of genetic information to build up specific enzymes for the biodegradation. Extracellular hydrolytic enzymes thus produced can disrupt major chemical bonds in the toxic molecules and results in the reduction of their toxicity (Sharma and Shah, 2005; Lal and Saxena, 1982). Microorganisms are the key factor in nutrient transformation. Microorganisms that involve in composting process excrete several extra-cellular enzymes include lignocelluloses, proteases, lipases and other enzymes that contribute in degradation of macro-molecules of organic wastes. Therefore inoculation of suitable microbial strains in initial organic wastes resulted in enrichment in the nutrient status of composts (Lei and Vander Gheynst, 2000).

Lipase are serine hydrolases of considerable physiological significance and industrial potential that can catalyze numerous reactions such as hydrolysis, inter-esterification, esterification, alcoholysis and aminolysis (Jaeger and Eggert, 2002). Biodegradation is breakdown of organic contaminants occurring due to production of extracellular enzymes by microorganisms. These contaminants can be considered as the substrate or microbial food source (Maier *et al.*, 2000). Biological treatment of municipal wastewater is successful when performed with activated sludge (Grady and Lim, 1980; Eckenfelder, 1985).

The microbial source of amylase is preferred to other sources because of its plasticity and vast availability. Microbial amylase has almost surpassed the synthetic sources in different industries (Pandey *et al.*, 2000). Amylases can be obtained from several sources such as plant, animal and microbes (Kathiresan and Manivannan, 2006). Starch degrading amylolytic enzymes are most important in the biotechnology industries with huge application in food, fermentation, textile and paper (Pandey *et al.*, 2000). Unusual bacterial amylases are found in acidophilic, alkalophilic and thermoacidophilic bacteria (Boyer and Ingle, 1972). In 2003, Windrow and aerated pile composting of sewage sludge and food waste could enhance the stability of these wastes and ensure the inactivation of pathogens and parasites (Furhacker and Haberl, 1995; Rantala *et al.* 2000; Krogmann, 2001).

FACTORS AFFECTING OF KITCHEN WASTE DEGRADATION

These factors are involved in the composting of kitchen wastes, which are discussed given below. Such as the composition of raw material, temperature, oxygen concentration, pH and water of the composting process-all influence the concentrations of gases produced (Wang *et al.*, 2002). The major advantage of aerobic composting over anaerobic composting is that it is fast and the decomposition process is completed within eight to twelve weeks and foul smelling gases are not produced. There are many factors that affect the composting process, such as the microbial diversity, proportions of the mixture, temperature, the aeration rate, oxygen consumption rates, compost recycling, moisture content, pH and C/N, and so on (Yanjun *et al.*, 2009; Boulter *et al.*, 2000).

THE COMPOSITION OF RAW MATERIAL

Where the amount of readily degradable compounds (sugars, carbohydrates, hemicelluloses and cellulose) is adequate, and the compost pile is well aerated and insulated, the pile temperature will generally rise within several days to the mesophilic and then to the thermophilic phase (40° to 70°C), reflecting a vigorous microbial activity and a rapid rate of organic compounds degradation. The degradation by thermophilic bacteria at the thermophilic phase is critical for pathogen control and killing of fly larvae, weed seeds if the high temperature is maintained three or more days. The composting rates decrease when the temperature reaches 60°C or above (Boch *et al.*, 1984). The temperature of the composting pile can be controlled by several strategies including configuration of the compost heap (its size and shape), by turning and watering, or by temperature feedback controlled ventilation (de Bertoldi *et al.*, 1985).

C/N RATIO

Carbon to nitrogen ratio of compost is the ratio between carbon and nitrogen content in compost. It is an important parameter of composting because its show about the compost quality. The composting may be more effective when C/N ratio between 30 to 40% (Chang and Chen, 2010). Some studies have shown that C/N ratio at lower than 20 is also effective (Kumar *et al.* 2010; Zhu, 2007). As aerobic composting is based on microbial respiration, so microorganisms

require food, energy and habitat. They need C as their energy source in order to manufacture enzymes for complex carbohydrates degradation into simpler forms which are then being used by these microorganisms (Hamdy, 2005). These nutrients are the basic building blocks of life and allow microbes to create the necessary enzymes which help them in the breakdown of contaminants. Carbon is the most basic element of living forms which is needed in greater quantities to microbes than other elements. Microbial cells constitute about 95% of hydrogen, oxygen, and nitrogen of the weight of cells. Phosphorous and sulfur contribute with 70%. The nutritional requirement of C/N ratio is 10:1, and carbon to phosphorous is 30:1. (Vidali, 2001).

TEMPERATURE

Temperature rises when sufficient heat is trapped within the compost pile. Composting can generally be divided into three stages based on the temperature in the composting pile. The rise in temperature and the rate of CO₂ release from the compost are directly related to O₂ consumption rate (Epstein, 1996). Actinomycetes tend to be common in the later stages of composting and can exhibit extensive growth. Bacteria are by far the most important decomposers during the most active stages of composting due to rapid growing ability on soluble substrates and tolerant of high temperatures. Thermophilic bacteria are dominant species at temperatures above 55°C, which kill pathogens (Strom, 1985).

According to Cooney and Emerson, (1964) thermophilic microorganisms require a maximum temperature for growth at or above 50°C and a minimum temperature for growth at or above 20°C. Thermophiles are reported to contain proteins which are thermostable and resist denaturation and proteolysis (Kumar and Nussinov, 2001). Specialized proteins known as chaperones are produced by these organisms, which help, after their denaturation to refold the proteins to their native form and restore their functions (Everly and Alberto, 2000). The cell membrane of thermophiles is made up of saturated fatty acids, which provides a hydrophobic environment for the cell and keeps the cell rigid enough to live at elevated temperatures (Herbert & Sharp, 1992).

MOISTURE

Moisture is also a very important parameter of composting it can maintain the rate of biodegradation of materials. The moisture content in excess amount can be negative for composting process. Bulking agents like sawdust, rice husk, cotton waste and maize straw can control the moisture content of composting materials (Adhikari et al., 2008; Paredes et al., 1996; Zorpas and Loizidou, 2008). When the moisture content between 55% to 65% the composting will more effective (Chang and Chen, 2010). Water is the media for nutrient transportation and metabolic reactions. The availability of nutrients and contaminants to microorganisms is affected by the water content in their micro-environment especially in the thin liquid layers on the surfaces of particles. The optimum moisture content is in the range of 50- 70%. If too much water fills the voids, the pore space that allows air diffusion would be limited (Gajalakshmi and Abbasi, 2008).

OXYGEN CONCENTRATION

The oxygen requirement for decomposition of food wastes was estimated to be 5.0 g O₂ g⁻¹ dry weight of food wastes. In this slurry-phase decomposition, suspended solids in the reactor disappeared at a maximum rate of 7.9 g dry weight L⁻¹ day⁻¹ (Yun et al., 2000). The use of aerobic thermophilic treatment to convert a mixture of sewage sludge and food waste into fertilizer has been reported previously (Wang et al., 2003a, b). Theoretically, the water content of the compost could reach 100% without causing harmful effects itself. However, as water content increases, the rate of O₂ diffusion decreases. As O₂ becomes insufficient to meet the metabolic demand, the composting process slows down and become anaerobic. The upper limit of water content is between 60 and 80%, depending on the composting materials (Golueke et al., 1991). Nakasaki et al., (1990) revealed that the supply of oxygen is essential for the escalation/ growth of composting process.

pH

The pH is a very important factor in composting and vermicomposting. It can increase and decrease the rate of biodegradation. Composting is a biological process of microbial activities and microbes survives at specific pH. Bulking agents like sawdust, cow dung and rice husk etc can control the pH value of composting (Chang and Chen, 2010; Adhikari et al., 2008; Singh and Kalamdhad, 2012). For the better biodegradation of composting materials the pH value should not very basic and very acidic it should be between the range of 6-8 (Chang and Chen, 2010; Gea et al., 2007; Alburquerque et al., 2006). Actinomycetes prefer moist but aerobic conditions with neutral or slightly alkaline pH. But reducing pH to reduce NH₃ volatilization by adjusting the pH of the compost to or below is possible, but low pH interferes with the transition of mesophilic degradation to thermophilic degradation (Sundberg et al., 2004). The

volatilization of organic acids (e.g., acetic acid), or odor, is reduced when the compost pile pH is raised to pH 5 or above (Brinton, 1997).

Although microbes are the real agents responsible for composting, their type and population size rarely are a limiting factor (Golueke, 1989). Bacilli predominated beyond the initial mesophilic phase. Compost turning or aeration is critical for a rapid degradation and high quality compost particularly for the food waste composting (Illmer and Schinner, 1997). A moisture level of 40 to 60% by weight should be maintained throughout the composting period (The U.S. Composting Council, 1997). Bacillus species are generally industrial importance since they have the ability to produce extracellular enzymes (proteases, amylases, cellulases, lipases, pectinases, and xylanases) that are active and stable at high pH values (Martens et al., 2001).

WATER

Water is a critical factor in composting system. Microbial cells have a physiological need for water. In addition, water can function as a solvent of substrates and salts, a major heat storage medium due to its high specific heat capacity, and as temperature adjusting substances through evaporation. Theoretically, the water content of the compost could reach 100% without causing harmful effects itself. However, as water content increases, the rate of O₂ diffusion decreases. As O₂ becomes insufficient to meet the metabolic demand, the composting process slows down and become anaerobic. The upper limit of water content is between 60 and 80%, depending on the composting materials (Golueke et al., 1991).

AERATION

Aeration or air supply is the major factor in the composting rate. It can highly affect the composting process and rate (Chang and Chen, 2010). Air supply should be proper in the compost materials for better and fast biodegradation of materials (Barrington et al., 2003). Food waste is an easily degradable material under aerobic conditions. A 91% reduction of food wastes was achieved during treatment in the slurry phase at 30°C with the addition of a mixture of microorganisms used in composting (Yun et al., 2000). The maximum content of microorganisms was 5 × 10¹⁰ cells ml⁻¹ at the 20th day (Park et al., 2002). The effect of aeration rate was considered to be a major factor affecting the slurry-phase decomposition of food wastes (Park et al., 2001).

In recent study, an attempt is made to enhance the aerobic composting through bio-stimulation by the addition of nutrients in order to enhance the efficacy of microorganisms (Martinez et al., 2008). Urea and tri super phosphate (TSP) are used as nutrient source. Indigenous microorganisms are stimulated by injecting these trace nutrients to waste system to increase the activity of microorganisms (Koenigsberg et al., 2005). In low-income countries, the solid waste generation rate is average only 0.4 to 0.6 kg/person/day, as opposed to 0.7 to 1.8 kg/person/day in fully industrialized countries (Cointreau, 1982; Blight and Mbande, 1996). In urban areas, solid waste has a very high organic content that ranged from 70- 85 %. Thus, under proper conditions this biodegradable fraction could be composed or co-composted for beneficial use as soil conditioners and bio-fertilizers (Parr and Hornick, 1992). The microbial population of soils is made up of five major groups including bacteria, actinomycetes, fungi, algae and protozoa, and among these groups, bacteria are the most abundant group (Alexander, 1961) and the most important microbe for decomposing waste.

Various composters are currently commercially available or several types of in-vessel composting systems have been developed for installation in food service establishments to manage food waste as a recyclable resource. It is difficult to maintain steady degradation due to the instability of the micro flora within the compost due to the raw material, pH, temperature and other environmental conditions (Miller, 1959; Summerell and Burgess, 1989). Composting is a microbial decomposition process in which easily degradable and putrescent organic waste materials are converted into a stable material, compost (Gray et al., 1971). It has been thought that the micro-organisms contributing to organic matter decomposition will change as composting progresses, since temperature, pH, moisture content, the quality and quantity of organic materials also change during composting (Gray et al., 1971; Golueke, 1977; Finstein and Morris, 1975; de Bertoldi et al., 1983).

Under anaerobic condition, microorganisms break down those materials into small molecules, produce organically reduced sulfur compounds, such as H₂S (Rapper and Muller, 2005), acids (Wu, et al., 2010), ammonia (Kim et al., 2009; Vijaya et al., 2010) and amines (R-NH₂) (Rapper and Muller, 2005) which are the most likely source of malodor (Rosenfeld et al., 2000). Bacteria use wastes for their own metabolism and finally they produce some simple and useful compounds which are important for soil health, plant growing and over all to keep well balance of natural ecosystem. Bacteria along with saprobic fungi are an important contributor to optimal agricultural and kitchen wastes bioconversion (Rahman et al., 2009). It contains mostly organic wastes which can be decomposed by composting process (Raimbault et al., 1998). During the process, part of organic C is released as CO₂, part incorporated into microbial cells and part humified. The organic nitrogen primarily as protein prior to composting is mineralized to inorganic N (NH₄-N and NO₃-N), which is then re-

synthesized into other forms of organic N in microbial biomass and humic substances during the composting process.

Degradation of organic C during composting is carried out by bacteria, fungi, and actinomycetes, depending on the stage of degradation, the characteristics of materials, and temperature (Epstein et al., 1996; USDA, 2000). There are many thermophilic actinomycetes, which can tolerate composting temperatures in the 50°C and low 60°C. In some countries, before 1980's, the soil fertility was maintained mainly by the use of organic fertilizers such as farmyard, manure, compost, green manure, straw and organic wastes. At the present, low soil organic matter content on arable lands has made them become less fertile (Yang et al., 1997).

In composting, a large number of microorganisms such as fungi, actinomycetes and bacteria, which at the same time attack the organic residues, use a portion of them to form their cellular material and produced carbon dioxide (Deporter et al., 1998). In addition, a variety of organic metabolites and water are produced. As the process reaches at its mature stage, a lower concentration of carbon dioxide is given off since microbial utilization of carbon source render the labile compounds less accessible due to diminution in the residues (Wang et al., 2004). Furthermore, composting is an economic, hygienic and ecological system of managing municipal as well as industrial solid wastes constituting biodegradable residues, both by aerobic and anaerobic process (Nakasaki et al., 1985). In general, composting is practiced in open piles or windrows (Gaur, 1982).

METHODS OF KITCHEN WASTE TREATMENT:-

COMPOSTING

An aerobic, biological process of degradation of biodegradable organic matter, Composting also leads to reduction in odour and the removal of pathogens. The wastes from wholesale fruit and vegetable markets, supermarkets and food processing are the best materials for composting (Sharma et al., 1997). Other reports describe the application of composts from different food wastes such as leaf compost (Maynard and Hill, 2000) and compost from agro-industrial wastes (Garcia-Gomez et al., 2002) which improve the physical properties of the soil and the produced crops. The optimal conditions for composting of food waste are: moisture content of the composting material must be at least 65%; pH near neutral; the C/N ratio of the material must be between 25: 1 and 35: 1. Additional aeration can improve the process; a temperature of 60°C must be kept for optimal thermophilic composting (Haug, 1993). Another aerobic technology for the treatment of food waste is a slurry technology which leads to a complete high rate decomposition of food waste to inorganic carbon (Park et al., 2001)

INCINERATION

A process of combustion designed to recover energy and reduce the volume of waste going to disposal. In its dry form, sewage sludge could be considered a special type of renewable fuel, due to the high quantity of organics of sufficiently high calorific value, similar to that of brown coal (Werther and Ogada, 1999; Spliethoff, 2010; Garcia, et al., 2013). There is therefore increased interest in utilisation of sewage sludge, resulting also from limited reserves of fossil fuels (and limited global security of energy supplies) and environmental and climatic regulations on CO₂ emissions. There are several thermal technologies for utilising municipal sewage sludge to obtain useful forms of energy, such as pyrolysis, gasification, combustion, and co-combustion processes (Manara and Zabaniotou, 2012; Calvo et al., 2013; Chun et al., 2011; Judex, et al., 2012; Rushidi et al., 2013). The advantages of thermal processes are the large reduction in volume, the thermal destruction of toxic organics and the recovery of the energy of organic sources in the sludge. One of the most promising options is combustion. Combustion characteristics of sewage sludge have been widely studied using thermo-analytical techniques. The influence of temperature and atmosphere has been investigated by thermal analytical techniques (Viana et al., 2011; Varol et al., 2010; Magdziarz and Wilk, 2013; Otero et al., 2007; Wu et al., 2012). Thermogravimetric analysis of the kinetics of combustion processes has been performed by Scott et al., 2006; Lee and Bae, 2009; Magdziarz, and Wilk, 2013). Landfill causes ground water contamination while in case of incineration; there is a problem of green house gas emissions, leading to the environmental degradation (Tsai et al., 2007; Lee et al., 2009).

LANDFILL

The deposition of waste in a specially designated area, which in modern sites consists of a pre-constructed 'cell' lined with an impermeable layer (man-made or natural) and with controls to minimize emissions (Rushton, 2003). High-quality compost is produced by interaction of many organisms that have suitable properties for the composting processes. Nevertheless, little information has been reported about in situ functions and roles of individual microbes in the composting processes, because many microbes related to composting are difficult to isolate and are characterized by conventional cultivation methods (Atkinson et al., 1996; Gautam et al., 2009).

RECYCLING

The recovery of materials from products after they have been used by consumers, Recycling of food wastes can be benefits to the environment by reducing the amount of garbage disposed, promoting the fertility of soil and improving the physical and chemical properties of soil (Park et al., 2002). Furthermore, recycling of food wastes reduce the unpleasant odors of garbage, benefits the sanitation of the environment, and decreases garbage collection-related spending. Food waste is less harmful to the environment than industrial waste. Thus, composting of food waste is attracting considerable attention because it would significantly reduce the amount of waste and the product can be used as compost or bio-fertilizer which can be handled, stored, transported and applied to the field without adversely affecting the environment (Debosz et al., 2002).

SEWAGE TREATMENT

Process of treating raw sewage to produce non-toxic liquid effluent which is discharged to rivers or sea and a semi-solid sludge, which is used as a soil amendment on land, incinerated or disposed in a landfill. Wastes include solids, liquids and gases, among them solid or semi solid forms are called "solid waste". Kitchen waste sewage (the liquid portion of kitchen food waste) contains carbohydrates, proteins, fats, and cellulose etc.; those are the source of nutrition for microbial metabolism (Wang et al., 2003). Kitchen waste has high moisture content (about 70%), malodor, high oil and grease levels as well as salt. If treated inadequately by land filling, the unpleasant odors, and ground-water pollution may cause residents to protest (Chen et al., 2008). This bio-reactor need not add any chemicals, operates easily, and almost every household can apply this cheap bio-reactor to treat their own kitchen sewage. This kind of household treatment can reduce kitchen waste weight about 40% for easier garbage truck collection. The cyanobacteria can decay organic pollutants and decreases water pollution (Cheunbarn and Peerapornpisal, 2010).

CONCLUSIONS

Based on the literature survey, it can be concluded that kitchen waste is very useful and very harmful to our day life in different conditions. It can be used for biogas, an organic acids, biofertilizers, and biomass production as humus which contain biological activity and associated with many major roles for fixation of environmental parts. Due to its degradation is becoming a great challenge because kitchen waste made up of biological polymer substances which provide nutritional substances for growing pathogenic microorganisms. Various treatment methods are available in the literature but the cheapest, eco-friendly and acceptable method is aerobic degradation (composting) by aerobic microbes. The anaerobic microbes release the extracellular enzymes such as hydrolytic enzymes involve in degradation of kitchen waste in anaerobic conditions and produces methane gas act as a global warming factor. So composting methods is a suitable for controlling of methane and global warming effect in environment.

REFERENCES

- ADHIKARI, B.K., BARRINGTON, S., MARTINEZ, J., KING, S. 2008. Characterization of food waste and bulking agents for composting. *Waste Management*, 28,795-804.
- ALBURQUERQUE, J., et al. 2006. Effects of bulking agent on the composting of "alperujo", the solid by-product of the two-phase centrifugation method for olive oil extraction. *Process Biochemistry*, 41, 127-132.
- ALEXANDER, M. 1961. Introduction to Soil Microbiology, *John Wiley and Sons, Inc*, New York.
- AOSHIMA, M., PEDRO, M.S., HARUTA, S., DING, L., FUKADA, T., KIGAWA, A., KODAMA, T., ISHII, M., IGARASHI, Y. 2001. Analyses of microbial community within a composter operated using household garbage with special reference to the addition of soybean oil. *Journal of Bioscience and Bioengineering*, 91, 456-461.
- ATKINSON, C.F., JONES, D.D., GAUTHIER, J.J. 1996. Biodegradability's and microbial activities during composting of oxidation ditch sludge. *Compost Science and Utilization*, 4, 84-96.
- BALASUNDARAN, M. 1999. Chemistry and Process of Composting. In: Balasundaran, M., Sharma, J.K., and Chacko, K.C., (eds.), *Compost for Container Seedling Production in Forest Nurseries*. Kerala Forest Research Institute, Peechi, Thrissur. pp 31-36.
- BARRINGTON, S., CHOINIÈRE, D., TRIGUI, M., KNIGHT, W. 2003. Compost convective airflow under passive aeration. *Bioresource Technology*, 86, 259-266.
- BLIGHT, G.E., MBANDE, C.M. 1996. Some problems of waste management in developing countries. *Journal of Solid Waste Technology and Management*, 23(1), 19-27.
- BOCH, P.D., SCHODA, M., KUBOTA, H. 1984. *Journal of Fermentation Technology*, 62, 285.

- BOULTER, J.I. BOLAND, G.J., TREVOR'S, J.T. 2000. Compost: A study of the development process and end-product potential for suppression of turf grass disease. *World Journal of Microbiology and Biotechnology*, 16, 115-134.
- BOYER, E.W., INGLE, M.B. 1972. Extracellular alkaline amylase from *Bacillus* sp. *Journal of Bacteriology*, 110, 992-1000.
- BRINTON, W.F. 1997. Proc. 6th annual conference on composting. October 11-13, Beltsville, MD.
- CALVO, A.I., et al. 2013. Particulate emissions from the co-combustion of forest biomass and sewage sludge in a bubbling fluidized bed reactor. *Fuel Processing Technology*, 114, 58-68.
- CEKMECELOGLUL, D., DEMIRCIL, A., GRAVESL, R.E., DAVITT, N.H. 2005. Applicability of optimised in-vessel food waste Composting for windrow systems. *Biosystems Engineering*, 91(4), 479-486.
- CHANDRAKANT, S. KARIGAR, AND SHWETHA, S. RAO. 2011. Role of Microbial Enzymes in the Bioremediation of Pollutants: *Enzyme Research*, 11, 1-11.
- CHANG, J.I., CHEN, Y.J. 2010. Effects of bulking agents on food waste composting. *Bioresource Technology*, 101, 5917-5924.
- CHANG, J.I., TSAI, J.J., WU, K.H. 2005a. Mathematical model for CO₂ evolution from the thermophilic composting of synthetic food wastes made of dog food. *Waste Management*, in press.
- CHEN, W.C., GENG, D.S., CHEN, W.C. 2008. The strategy and bioenergy potential for kitchen waste recycling in Taiwan. *Journal of Environmental Engineering and Management*, 18(4), 281-287.
- CHEUNBARN, S., PEERAPORNPIHAL, Y. 2010. Cultivation of *Spirulina platensis* using anaerobically swine wastewater treatment effluent. *International Journal of Agriculture and Biology*, 12.
- CHOI, M.H., PARK, Y.H. 1998. The influence of yeast on thermophilic composting of food waste. *Letters in Applied Microbiology*, 26, 175-178.
- CHUN, Y.N., KIM, S.C., YOSHIKAWA, K. 2011. Pyrolysis gasification of dried sewage sludge in a combined screw and rotary kiln gasifier. *Applied Energy*, 88, 1105-1112.
- COINTREAU, S. 1982. Environmental management of urban solid wastes in developing countries, A Project Guide. *Urban Development, World Bank*, Washington, DC.
- COONEY, D.G., EMERSON, R. 1964. Thermophilic Fungi: An Account of their Biology, Activities and Classification. W. H. Freeman and Co., San Fransisco.
- CURTIES, S.E. 1981. Environmental Management in Animal Agriculture, Animal Environment Services Mahomet III, pp. 24-27.
- DAS, K.C., TOLLNER, E.W. EITEMAN, M.A. 2003. Comparison of synthetic and natural bulking agents in food waste composting. *Compost Science & Utilization*, 11(1), 27-35.
- DE BERTOLDI, M., VALLINI, G., PERA, A. 1983. The biology of composting: a review. *Waste Management & Research*, 1, 157-176.
- DE BERTOLDI, M., VALLINI, G., PERA, A. 1985. In: Gasser, J. K. R. (Ed.). Technological aspects of composting including modeling and microbiology. Composting of agricultural wastes and other wastes. *Elsevier Applied Science Publication*, NY.
- DEBOSZ, K., PETERSEN, S.O., KUBE, L.K., AMBUS, P. 2002. Evaluating effects of sewage sludge and household compost on soil physical, chemical and microbiological properties. *Applied Soil Ecology*, 19, 237-48.
- DEPORTER, I., BENOIT-GUYOD, J.L., ZMIROU, D., BOUVIER, M. 1998. Microbial Disinfection capacity of municipal solid waste water composting. *Journal of Applied Microbiology*, 85, 238-246.
- DONAHUE, D.W., CHALMERS, J.A. STOREY, J.A. 1998. Evaluation of in-vessel composting of university postconsumer food wastes. *Compost Science and Utilization*, 6(2), 75-81.
- DROFFNER, M.L., BRINTON, W.F. 1995. Survival of *E. coli* and *Salmonella* population in aerobic thermophilic composts as measured with DNA gene probes. *Zentralbl Hygiene*, 197, 387-397.
- ELEWELL, D.L., KEENER, H.M. HANSEN, R.C. 1996. Controlled high rate composting of mixtures of food residuals, yard trimmings and chicken manure. *Compost Science and Utilization*, 4(1), 6-15.
- ENEJI, A.E., YAMAMOTO, S., HONNA, T., ISHIGURO, A. 2001. Physicochemical changes in livestock feces during composting. *Communication in Soil Science and Plant Analysis*, 32, 477- 489.
- EPSTEIN, E. 1996. The Science of composting. *Technomic Publishing Company, Inc.*, Lancaster, PA.
- ECKENFELDER, W.W., PATOCZKA, J., WATKIN, A. T. 1985. *Chemical Engineering*, 90, 60.
- EVERLY, C., ALBERTO, J. 2000. Stressors, stress and survival: overview. *Frontiers in Bioscience*, 5, 780-786.
- FAUCETTE, B., DAS, K.C., RISSE, M. 2001. University tests in-vessel composting of food residuals. *Biocycle*, 42(1), 68-70.
- FILIPPI, C., BEDINI, S., LEVI-MINZI, R., CARDELLI, R., SAVIOZZI, A. 2002. Co-composting of olive oil mill by-products: chemical and microbiological evaluations. *Compost Science and Utilization*, 10(1), 63-71.
- FINSTEIN, S.M., MORRIS, L.M. 1975. Microbiology of Municipal Solid Waste composting. *Advanced Applied Microbiology*, 19, 113-151.
- FUHACKER, M., HABERL, R. 1995. Composting of sewage sludge in a rotating vessel. *Water Science and Technology*, 32, 121-125.
- GAJALAKSHMI, S., ABBASI, S.A. 2008. Solid waste management by composting: State of the art. *Critical Reviews in Environmental Science & Technology*, 38, 311-400.
- GARCIA, G., ARUAZO, J., GONZALO, A., SANCHEZ, J.L., ABREGO, J. 2013. Influence of feedstock composition in fluidized bed co-gasification of mixtures of lignite, bituminous coal and sewage sludge. *Chemical Engineering Journal*, 222, 345-352.
- GARCIA-GOMEZ, A., BERNAL, M.P., ROIG, A. 2002. Growth of ornamental plants in two composts prepared from agro-industrial wastes. *Bioresource Technology*, 83, 81-87.
- GAUR, A.C. 1982. A Manual of Rural Composting. Field Document No, 15, FAO, Rome, pp. 102.
- GAUTAM, S.P., ET AL. 2009. Biodegradation and Recycling of Urban Solid Waste. *American Journal of Environmental Science*, 5(5), 553-556.
- GEA, T., BARRENA, R., ARTOL, A., SANCHEZ, A. 2007. Optimal bulking agent particle size and usage for heat retention and disinfection in domestic wastewater sludge composting. *Waste Management*, 27, 1108-1116.
- GOLUEKE, C.G. 1989. The bicycle guide to the art and science of composting. *The JG Press*, Emmaus, PA.
- GOLUEKE, C.G. (1977). Biological Reclamation of Solid Wastes. *Rodale Press, Inc*, Emmaus, PA, USA.
- GOLUEKE, C.G., (1991). *Bio-cycle*, 31(9), 70.
- GRADY, C.P.L., LIM, H.C. 1980. Biological wastewater treatment- theory and applications, Marcel Dekker, (Ed.), New York, Basel.
- GRAY, K.R., SHERMAN, K., BIDDLESTONE, A.J. 1971. A review of composting- Part I. *Process Biochemistry*, 6, 22-28.
- GUPTA, S., KRISHNA, M., PASAD, R.K., GUPTA, S., KANSAL, A., 1998. Solid waste management in India: options and opportunities. *Resource Conservation Recycle*, 24, 137-154.
- HAMDY, H.S. 2005. Purification and characterization of the pectin lyase produced by *Rhizopus oryzae* grown on orange peel. *Annals of Microbiology*, 55, 205-21.
- HAO. X.Y., CHAG, C., LARNEY, F.J., TRAVIS, G.R. 2001. *Journal of Environmental Quality*, 30, 376.
- HAUG, R.T. 1993. The Practical Handbook of Compost Engineering. *Lewis Publishers*, Boca Raton, FL, USA.
- HERBERT, R., SHARP, R. 1992. Molecular biology & biotechnology of extremophiles. *Chapman and Hall*, New York.
- HOORNWEG, D., THOMAS, L. 1999. What a Waste: Solid Waste Management in Asia. Washington, DC, USA.
- ILLMER, P., SCHINNER, F.N 1997. *Bioresource Technology*, 59,157.
- JAEGER, K.E., EGGERT, T. 2002. Lipases for biotechnology. *Current Opinion in Biotechnology*, 13, 390-397.
- JEAN, NATHALIE -BAPTISTE, 2009. People & Food waste-The practice of everyday life;
- JOHRI, B.N., SATYANARAYANA, T., Olsen, J., 1999. Thermophilic moulds in Biotechnology. *Kluwer Academic Publishers*, USA.
- JUDEX, J.W., GAIFFI, M., BURGBACHER, H.C. 2012. Gasification of dried sewage sludge: Status of the demonstration and the pilot plant. *Waste Management*, 32, 719-723.
- KASHYAP, D.R., DADHICH, K.S., SHARMA, S.K. 2003. Biomethanation under psychrophilic conditions: a review. *Bioresource Technology*, 87, 147-153.
- KATHIRESAN, K., MANIVANAN, S. 2006. Alpha-amylase production by *Penicillium fellutanum* isolated from mangrove rhizosphere soil. *African Journal of Biotechnology*, 5, 829-832.
- KIM, K.H., PAL, R., AHN, J.W., KIM, Y.H. 2009. Food decay and offensive odorants: a comparative analysis among three types of food. *Waste Management*, 29(4) 1265-73.
- KOENIGSBERG, S.S., HAZEN, T.C. PEACOCK, A.D. 2005. Environmental Biotechnology: A Bioremediation Perspective. *Remediation*, 15, 5-25.
- KROGMANN, U. 2001. Composting. In *Sludge into Bio-solids: Processing, Disposal, Utilization* ed. Spinosa, L. and Vesilind, P.A. pp. 259-277. London: IWA Publishing.
- KUMAR M., OU, Y.L., LIN, J.G. 2010. Co-composting of green waste and food waste at low C/N ratio. *Waste Management*, 30, 602-609.
- KUMAR, S., NUSSINOV, R. 2001. How do thermophilic proteins deal with heat? A review. *Cellular and Molecular Life Sciences*, 58, 1216-1233.
- LAL, R., SAXENA, D.M. 1982. Accumulation, metabolism & effects of organochlorine insecticides on microorganisms, *Microbiology Review*, 46, 95-127.
- LAOS, F., MAZZARINO, M.J., WALTER, I., ROSELLI, L. 1998. Composting of fish waste with wood by-products and testing compost quality as a soil amendment: experiences in the Patagonia Region of Argentina. *Compost Science and Utilization*, 6(1), 59-66.
- LEE, D.H., BEHERA, S.K., KIM, J.W., PARK, H.S. 2009. Methane production potential of leachate generated from Korean food waste recycling facilities: A lab-scale study. *Waste Management*, 29, 876-882.

- LEE, H.S., BAE, S.K. 2009. Combustion kinetics of sewage sludge and combustible wastes. *Journal of Material Cycles and Waste Management*, 11, 203–207.
- LEI, F., VANDER GHEYNST, J.S. 2000. The effect of microbial inoculation and pH on microbial community structure changes during composting. *Process Biochemistry*, vol. 35, pp. 923–929.
- LIANG, Y. 2000. Nitrogen Retention in the High Stage of Composting. Ph.D. Dissertation, University of Alberta, Edmonton, Alberta, Canada.
- MAC DONALD DOW, M.G.C., GRIFFIN SHAY, E. 1981. Returning wastes to the land. Report of the advisory committee on technology for international development commission on international relations. *National Research Council, National Academy Press*, Washington, DC.
- MAIER, R.M., PEPPER, I.L., GERBA, C.P. 2000. *Environment Microbiology*, Academic Press.
- MAGDZIARZ, A., WILK, M. 2013. Thermogravimetric study of biomass, sewage sludge and coal combustion. *Energy Conversion and Management*, 75, 425–430.
- MANARA, P., ZABANIOTOU, A. 2012. Towards sewage sludge based biofuels via thermo chemical conversion – a review. *Renew. Sustain. Energy Review*, 16, 1081–1087.
- MARTINEZ, A.S., CORDOVA, M.S., CRUZ, O.S., DELGADO, E., ANDRADE, H.P., MARROQUIN, L.A.H. ROLDAN, H.M. 2008. Development of a bioremediation process by bio-stimulation of native microbial consortium through the heap leaching technique. *Journal of Environmental Management*, 88, 115–119.
- MARTENS, W., MARTINEC, M., ZAPIRAIN, R., STARK, M., HARTUNG, E., PALMGREN, U. 2001. Reduction potential of microbial, odour and ammonia emissions from a pig facility by biofilters. *International Journal of Hygiene and Environmental Health*, 203, 335–345.
- MAYNARD, A.A., HILL, D.E. 2000. Cumulative effect of leaf compost on yield and size distribution in onions. *Compost Science and Utilization*, 8, 12–18.
- MCGUCKIN, R.L., EITEMAN, M.A., DAS, K.C. 1999. *Journal of Agricultural Engineering Research*, 72, 375–384.
- MILLER, G.L. 1959. Use of dinitrosalicylic acid reagent for determination of reducing sugars. *Analytical Chemistry*, 31, 426–428.
- MOHAREB, A.K., WARITH, M., NARBAITZ, R.M. 2004. Strategies for the municipal solid waste sector to assist Canada in meeting its Kyoto Protocol commitments. *Environmental Reviews*, 12, 71–95.
- NAKASAKI, K., HIDEKI, Y., YASUSHI, S., HIROSHI, K. 1990. Effect of oxygen concentration on composting of garbage. *Journal of Fermentation and Bioengineering*, 70(6), 431–433.
- NAKASAKI, K., SASAKI, M.S. KUBOTA, H. 1985. Effect of seeding during thermophilic composting of sewage sludge. *Applied and Environmental Microbiology*, 49, 724–726.
- OBIRE, O., NWAUBETA, O., ADUE, S.B.N. 2002. Microbial Community of a Waste- Dump Site. *Journal of Applied Sciences & Environmental Management*, pp. 78–83.
- OGBONNA, D.N., IGBENIJE, M. 2006. Characteristics of Microorganisms Associated with Waste Collection Sites in Port-Harcourt City, Nigeria. *Nigerian Journal of Microbiology*, 20(3), 1427–1434.
- OTERO, M., GOMEZ, X., GARCIA, A.I., MORAN, A. 2007. Effects of sewage sludge blending on the coal combustion: A thermogravimetric assessment. *Chemosphere*, 69, 1740–1750.
- PAGANS, E., BARRENA, R., FONT, X., SANCHEZ, A. 2006. NH₃ emissions from the composting of different organic wastes. *Chemosphere*, 62, 1534–1542.
- PAGANS, E., FONT, X., SANCHEZ, A. 2005. Biofiltration for ammonia removal from composting exhaust gases. *Chemical Engineering Journal*, 113, 105–110.
- PANDEY, A., NIGARM, P., SOCCOL, C.E., SOCCOL, V.T., SINGH, D., MOHAN, R. 2000. Advances in microbial amylases. *Biotechnology and Applied Biochemistry*, 31, 135–152.
- PAREDES, C., BERNAL, M.P., ROIG, A., CEGARRA, J., SANCHEZ-MONEDER, M.A. 1996. Influence of the bulking agent on the degradation of olive-mill wastewater sludge during composting. *International Biodeterioration & Biodegradation*, 38, 205–210.
- PARK, J.I., YUN, Y.S., PARK, J.M. 2001. Oxygen-limited decomposition of food wastes in a slurry bioreactor. *Journal of Industrial Microbiology and Biotechnology*, 27, 67–71.
- PARK, J.I., YUN, Y.S., PARK, J.M. 2002. Long-term operation of slurry bioreactor for decomposition of food wastes. *Bioresource Technology*, 84, 101–104.
- PARR, J.F. HORNICK, S.B. 1992. Utilization of municipal wastes, In: *Soil Microbial Ecology. Applications in agricultural and environmental management*, pp. 545–559. F.B. Metting (ed.) Marcel Dekker, Inc., New York, U.S.A.
- PARR, T.W., FERRETTI, M., SIMPSON, I.C., FORSIUS, M., KOVACS-LANG, E. 2002. Towards a long-term integrated monitoring programme in Europe: Network design in theory and practice. *Environmental Monitoring and Assessment*, 78, 253–290.
- PIKE, E.B. 1975. Aerobic Bacteria, in: *Ecological Aspects of Used Water Treatment*, Curds, C. R. and Hawkes, H. A. (Ed.), Vol. 1, *Academic Press*, London, pp 1–63.
- RAHMAN, M.M., BEGUM, M.F., KHAN, M.S.I., ET AL. 2009. Isolation, identification and cultural optimization of native bacteria isolates as a potential bioconversion agent. *Journal of Applied Science Research*, 5, 1652–1662.
- RAIMBAULT, M. 1998. General and microbiological aspects of solid substrate fermentation. *Electronic Journal of Biotechnology*, 1, 1–15.
- RANTALA, P.R., VAAJASAARI, K., JUVONEN, R., SCHULTZ, E., JOUTTI, A. MARELA-KURTTO, R. 2000. Composting and utilization of forest industry wastewater sludges. *Water Science and Technology*, 42, 227–234.
- RAPPER, S., MULLER, R. 2005. Microbial degradation of selected odorous substances. *Waste Management*, 25, 940–954.
- RATHI, S. 2006. Alternative approaches for better municipal solid waste management in Mumbai, India. *Journal of Waste Management*, 26, 1192–1200.
- RAYNAL, J., DELGENKS, J.P., MOLETTA, R. 1998. Two phase anaerobic digestion of solid wastes by a multiple liquefaction reactors process. *Bioresource Technology*, 65, 97–103.
- ROSENFELD, P.E., HENRY, C.L., DILLS, R.L., HARRISON, R.B. 2000. Comparison of odor emission from three different biosolids applied to forest soil. *Water, Air, & Soil Pollution*, 127, (1-4), 173–191.
- RUSHIDI, A.I., AL-MUTLAQ, K.F., SASMAL, S.K., SIMONEIT, B.R.T. 2013. Alteration of sewage sludge biomass into oil-like products by hydrous pyrolysis methods. *Fuel*, 103, 970–979.
- RUSHTON, L. 2003. Health hazards and waste management. *British Medical Bulletin*, 68, 183–197.
- SAEBO, A., FERRINI, F. 2006. The use of compost in urban green areas - a review for practical application. *Urban For. Urban Greening*, 4, 159–169.
- SAKAI, K., MURATA, Y., YAMAZUMI, H., TAU, Y., MORI, M., MORIGUCHI, M., SHIRAI, Y. 2000. *Food Science and Technology Research*, 6, 140.
- SARAVANNAN, S., MEENAMBAL, T., UMA, R.N. 2003. Study on biodegradation of fruit waste aerobic composting. Proceedings of the Third International Conference on Environment and Health, Chennai, India.
- SCOTT, S.A., DENNIS, J.S., DAVIDSON, J.F., HAYHURST, A.N. 2006. Thermogravimetric measurements of the kinetics of pyrolysis of dried sewage sludge. *Fuel*, 85, 1248–1253.
- SEYMOUR, R.M., DONAHUE, D., BOURDON, M., EVANS, J.R., WENTWORTH, D. 2001. Intermittent aeration for in-vessel composting of crab processing waste. *Compost Science and Utilization*, 9(2), 98–106.
- SHARMA, S., SHAH, K.W. 2005. Generation and disposal of solid waste in Hoshangabad. In: *Book of Proceedings of the Second International Congress of Chemistry and Environment*, Indore, India, pp. 749–751
- SHARMA, V.K., CANDIETELLI, M., FORTUNA, F., CORNACCHIA, G. 1997. Process of urban and agro-industrial residues by aerobic composting: review. *Energy Conversion and Management*, 38, 453–478.
- SINGH, J., KALAMDHAD, A.S. 2012. Concentration and speciation of heavy metals during water hyacinth composting. *Bioresource Technology*, 124, 169–179.
- SPLIETHOFF, H. 2010. *Power Generation from Solid Fuels*. Springer, Germany.
- STENTIFORD, E.J. 1987. Recent developments in composting. In: STROM, P.F. 1985. *Applied and Environmental Microbiology*, 50, 899.
- SUMMERELL, B.A., BURGOSS, L.W. 1989. Decomposition and chemical composition of cereal straw. *Soil Biology and Biochemistry*, 21, 551–559.
- SUNDBERG, C., SMARS, S., JONSSON, H. 2004. *Bioresource Technology*, 95, 145.
- SVENSSON, K., ODLARE, M., PELL, M. 2004. *Journal of Agriculture Science*, 142, 461–467.
- TAYLOR, D. J., GREEN, N.P.O., STOUT, G.W. 2002. *Biological Science*, 3rd Edition. Cambridge University Press. 984, 345.
- THE U.S. COMPOSTING COUNCIL. 1997. *Compost enhancement guide*. Alexandria, VA.
- TOMATI, U., MADEJON, E., GALLI, E., CAPITANI, D., SEGRE, A.L. 2001. Structural changes of humic acids during olive mill pomace composting. *Compost Science and Utilization*, 9(2), 134–142.
- TSAI, S.H., WEI, C.B., YANG, S.S. 2002. Quality of food waste compost produced by local autonomy group in Taipei City. *Journal of Biomass Energy Soc China*, 21, 103–117.
- TSAI, S.H., LIU, C.P. YANG, S.S. 2007. Microbial conversion of food wastes for biofertilizer production with thermophilic lipolytic microbes. *Renewable Energy*, 32, 904–915.
- USDA. 2000. *Composting*. Part 637, National Engineering Handbook, NRCS, U.S. Department of Agriculture, Washington, D.C.
- VAROL, M., ATIMTAY, A.T., BAY B., OLGUN, H. 2010. Investigation of co-combustion characteristics of low quality lignite coals and biomass with thermogravimetric analysis. *Thermochimica Acta*, 510, 195–201.
- VERMA, N., BANSAL, N.C., KUMAR, V. 2011. Pea peel waste: a lignocellulosic waste and its utility in cellulose production by *Trichoderma reesei* under solid state cultivation. *Bioresources*, 6, 1505–1519.

- VERRIER, D., RAY, F., ALBAGNAC, G. 1987. Two phase methanation of solid vegetable wastes. *Biological Wastes*, 22, 163–177.
- VIANA, M.M., MELCHERT, M.B.M., DE MORAIS, L.C., BUCHLEER, P.M., DWECK, J. 2011. Sewage sludge coke estimation using thermal analysis. *Journal of Thermal Analysis and Calorimetry*, 106, 437–443.
- VIDALI, M. 2001. Bioremediation. An overview. *Pure and Applied Chemistry*, 73, 1163-1172.
- VIJAYA, T., CHANDRA, M.K., DURGA, S.M.S., FAREEDA, G. 2010. Comparative studies on growth and remediation of waste water by two cyanobacterial biofertilizers. *Agriculture Conspectus Scientificus*, 75 (3), 99-103.
- VILLAGE, P.K.S. 1998. Characterization of Municipal Solid Waste in the United State. 1997 update. *US Environmental Protection Agency Municipal and Industrial Solid Waste Division*.
- WAI- OGOSU, O.A. 2004. Monitoring and Evaluation of Industrial Waste Management Options in Rivers State. Paper presented at a workshop on *Sustainable Environmental Practices in Rivers State* organized by Rivers State Ministry of Environment, Hotel Presidential, Port Harcourt, 23-24 March, 2004.
- WANG, J.Y., STABNIKOVA, O., IVANOV, V., TAY, S.T.L., TAY, J.H. 2003a. Intensive bioconversion of sewage sludge and food waste by *Bacillus thermoamylovorans*. *World Journal of Microbiology and Biotechnology*, 19, 427–432.
- WANG, J.Y., STABNIKOVA, O., IVANOV, V., TAY, S.T.L., TAY, J.H. 2003b. Intensive aerobic bioconversion of sewage sludge and food waste into fertilizer. *Waste Management and Research*, 21, 405–415.
- WANG, P., CHANGA, C.M., WATSON, M.E., DICK, W.A., CHEN, Y., HOITINK, H.A.J. 2004. Maturity indices for composted dairy and pig manures. *Soil Biology and Biochemistry*, 36, 767-776.
- WANG, Q., NARITA, J.Y., REN, N., FUKUSHIMA, T.Y., OHSUMI, Y., KUSANO, K., SHIRAI, Y., OGAWA, H.I. 2003. Effect of pH Adjustment on Preservation of Kitchen Waste Used for Producing Lactic Acid. *Water, Air, & Soil Pollution*, 144 (1-4) 405-418.
- WANG, Q.H., NARITA, J.Y., MORISHITA, M., OHSUMI, Y., KUSANO, K., SHIRAI, Y., OGAWA, H.I. 2001. *Process Biochemistry*, 37, 351.
- WANG, Q.H., NARITA, J.Y., REN, N.Q., FUKUSHIMA, T., OHSUMI, Y., KUSANO, K., SHIRAI, Y., OGAWA, H. 2002. *Water, air and soil pollution*, 144, 405
- WANG, Q.H., XU, Z., MENG, L.H., ET AL. 2003b. Influence of temperature on production of lactic acid from kitchen waste garbage. *Journal of Harbin Institute and Technology*, 10, 195-199.
- WERTHER, J., OGADA, T. 1999. Sewage sludge combustion. *Progress in Energy and Combustion Science*, 25, 55–116.
- WU, H., HANNA, M.A., JONES, D.D. 2012. Thermogravimetric characterization of dairy manure as pyrolysis and combustion feedstocks. *Waste Management and Research*, 30 (10), 1066–1071.
- WU, T., WANG, X., LI, D., YI, Z. 2010. Emission of volatile organic sulfur compounds (VOSCs) during aerobic decomposition of food wastes. *Atmospheric Environment*, 44, 5065-5071.
- YANJUN, L., XINGWU, W., JIFENG, G. 2009. Characteristics of municipal solid waste and sewage sludge co-composting. *Waste Management*, pp. 1152-1157.
- YANG, H.S., HANSEN, B.H. 1997. *European Journal of Agronomy*, 7, 211.
- YUN, Y.S., PARK, J.I., SUH, M.S., PARK, J.M. 2000. Treatment of food wastes using slurry-phase decomposition. *Bioresource Technology*, 73, 21–27.
- ZORPAS, A.A., LOIZIDOU, M. 2008. Sawdust and natural zeolite as a bulking agent for improving quality of a composting product from anaerobically stabilized sewage sludge. *Bioresource Technology*, 99, 7545-7552.
- ZHU, N. 2007. Effect of low initial C/N ratio on aerobic composting of swine manure with rice straw. *Bioresource Technology*, 98, 9-13.